Development of a Water Quality Model for the Queensland Murray-Darling Basin using the eWater Source Dynamic SedNet Framework

Prepared by:

Mr Wes Davidson

Department of Natural Resource, Mines and Energy, Toowoomba

© State of Queensland (Department of Natural Resource, Mines and Energy) 2018

Contact:

David Waters

Principal Scientist (modelling)

Department of Natural Resource and Mines

Email: David.Waters@dnrme.qld.gov.au

Phone: (07) 4529 1395

To reference this volume:

Davidson, W. (2018) Development of a Water Quality Model for the Queensland Murray-Darling Basin using the eWater Source – Dynamic SedNet Framework, Technical Report , Queensland Department of Natural Resources, Mines and Energy, Toowoomba, QLD

This document has been prepared with all due diligence and care, based on the best available information at the time of publication. The department holds no responsibility for any errors or omissions within this document. Any decisions made by other parties based on this document are solely the responsibility of those parties. Information contained in this document is from a number of sources and, as such, does not necessarily represent government or departmental policy.

Acknowledgments

Many thanks to those who contributed including:

- \triangleright Dave Waters, Robin Ellis and Shawn Darr for their assistance in collation and preparation of the data layers, and enabling me to gain a working knowledge of the Dynamic SedNet Catchment framework
- Tony King for his assistance with data collation and Jade Gould for data transformation.
- \triangleright Andrew Biggs and Dave Burton for their assistance with data layer collation and presentation
- \triangleright Scott McKie and Jade Gould for their support with water quality data collection

EXECUTIVE SUMMARY

Source Catchments Water Quality models were developed for the Queensland section of the Murray Darling Basin. The Queensland Murray Darling Basin (QMDB) Water Quality Models were built to assist in the development of water quality guidelines for Murray Darling Basin planning requirements. Total Suspended Sediment (TSS), Total Nitrogen (TN) and Total Phosphorus (TP) were the constituents of interest.

This work built on the experience of the Great Barrier Reef (GBR) Paddock to Reef Modelling Program and applied a similar modelling approach. Three separate models were created for the QMDB covering: South West catchments –namely the Bulloo, Paroo, Warrego and Nebine catchments (SWNRM), Condamine, Balonne and Maranoa catchments (CBM) and Moonie and Border Rivers catchments (MNBD). Models were calibrated using the Sacramento rainfall runoff model coupled to the Parameter Estimation Software Tool (PEST). Due to the limited water quality monitoring data available across the region for calibration, historical water quality (WQ) data was correlated against log transformed flow to build a relationship between TSS, TP and TN and gauge discharge by catchment. The resulting concentration values were used to calculate daily through to average annual loads. These loads were then used for model calibration.

The hydrological calibration achieved a percent bias (PBIAS) of less than 5% for 36 of the 37 gauges used for calibration for the 36 year modelling period. Modelled average annual TSS export loads for the 36 year modelling period were estimated to be 1,906 kt/yr for the SWNRM catchments, 198 kt/yr for CBM and 53 kt/yr for the MNBD catchments for the 36 year climate period (1980-2015). In terms of the overall QMDB sediment budget, gully erosion contributed 43%, streambank 37% and hillslope erosion 20% of the total sediment load exported. Limited measured data was available across the full range of flow heights for water quality calibration which meant that there is a degree of uncertainty about the measured estimates, a common problem worldwide.

The objective of the project was to develop a catchment model using the most up to date approach and data sets. This was achieved and the model has been used as one line of evidence in the development of high and low flow water quality guidelines for Water Quality Objectives for the Queensland Environmental Protection Policy. Additional water quality data collection at end of system gauges during high flow events will significantly improve model load estimates.

The model could be used and refined by regional Natural Resource Management (NRM) bodies in future years for scenario analysis such as prioritising natural resource management investment programs for improved land management practices. Using a model in a data poor area has highlighted the value of event monitoring to calibrate and validate water quality models. Development of such a model incorporating a range of erosion processes provided a basis for prioritising future research in catchments, in particular improve our understanding of sediment transport where limited measured data is available.

TABLE OF CONTENTS

LIST OF FIGURES

LIST OF TABLES

x

1 INTRODUCTION

The Commonwealth *Water Act 2007* provides instruments for a coordinated approach to the management of water resources throughout a Basin. The *Murray Darling Basin Plan (2012)* (Basin Plan) (MDBA website, 2017) is one such instrument prepared under Part 2 of the *Water Act 2007* (Commonwealth). The Basin Plan provides the framework for a coordinated approach across the five Basin States and Territories, for the management of water resources in the Murray-Darling Basin.

Basin Plan includes a range of management objectives and outcomes to be achieved in the Murray-Darling Basin in relation to water quality, with Chapter 9 stating the requirements of the water quality and salinity management plan; and Chapter 10, Part 7 stating the need for a water resource plan (WRP) to include a water quality management plan (WQM Plan).

*Water quality and salinity objectives, water quality targets for planning of water flows, water quality targets that apply to the preparation of the water resource plans, and water quality targets for the purposes of long-term salinity planning and management (item 10 of the table in subsection 22(1) of the Act). This Chapter also includes the key causes of water quality degradation in the Murray-Darling Basin. (*Basin Plan, Part 2, 1.05(1) Chapter 9)

As a basin State, Queensland supports the development of WQM Plans for inclusion in the WRP for each respective basin within the Queensland portion of the Murray Darling Basin.

With regard to the national framework, Queensland Department of Environment and Science also provides for the protection of water resources through the *Environmental Protection Act 1994* and the subordinate legislation *Environmental Protection (Water) Policy 2009 (EPP Water).* EPP Water provides a mechanism for Healthy Waters Management Plans (HWMPs) to be developed with the intent of improving water quality in a basin area. HWMP's include environmental values, water quality objectives and management goals for a basin area.

This project aims to develop a water quality model to support the development of HWMPs/WQM Plans, informing management actions and measures to achieve objectives as part of the respective Commonwealth accredited Water Resource Plans(WRPs). Basin and catchment scale water quality modelling is an essential tool that improves our understanding of water quality transport processes (sediment, nutrients, etc.). The model has the potential to be used by NRM groups to assist in prioritising on ground investment in improved land management practices.

This project will utilise the nationally adopted modelling framework (eWater Source) endorsed by both state and federal governments in July 2012. This project will also capitalise on learnings from the Reef Water Quality Protection Plan Water Quality Modelling (Waters et al., 2014) by applying its methodology where appropriate, to the Queensland Murray-Darling Basin (QMBD).

2 OBJECTIVES

The objective of the project are:

- Collate water quality data across the basin to derive relationship between flows and major water quality parameters (Total suspended sediment TSS, and nutrients) to calibrate the model.
- to develop a calibrated Water Quality Model incorporating functionality as used in the GBR modelling for Queensland Murray Darling Basin (QMDB) catchments
- Provide best estimates of sediment and nutrient export loads for all QMDB Basins
- Use the modelled outputs to assist in the development of water quality targets
- Consult with key stakeholder groups across the QMDB and Bulloo catchments to generate awareness of the water quality model and its potential application (Refer Table 34 for groups consulted).

3 HISTORY OF CATCHMENT WATER QUALITY MODELS IN THE QUEENSLAND MURRAY DARLING BASIN

There have been several water quality catchment models developed within the QMDB area. These have been created to support local Natural Resource Management (NRM) groups in testing on-ground investment strategies with the support of the Department of Natural Resources, Mines and Energy (DNRME). [Table 1](#page-11-1) below provides an overview of these models. These projects created a specific model for the area within the QMBD which was the target for the project using predecessors of the eWater Source Catchments platform. This project builds on the learnings from these previous models.

TABLE 1 - HISTORY OF CATCHMENT MODELS FOR QUEENSLAND MURRAY DARLING BASIN

3.1 PREVIOUS APPROACHES

These models were created to best represent the water quality processes within the catchments, using the best available data and software at the time. A summary of the models and parameters used are shown in [Table 2.](#page-12-1) In the development of the current model parameters, the previous approaches were taken into account.

TABLE 2 - PREVIOUS MODEL SETUP DETAILS

I

 3.2 IMPROVEMENT RECOMMENDATIONS

A list of recommendations from these projects are summarised i[n Table 3.](#page-13-4)

3.2.1 COMPLEMENTARY EVENT SAMPLING PROJECT

In each of the previous modelling projects, it was apparent that there was limited event data, across the range of flows, to calibrate or validate the models. An improvement in the collection of event data means a better calibration for high flows for constituents. Therefore, a separate project was established concurrently with this modelling project to use turbidity and electrical conductivity probes for continuous water quality monitoring combined with event sampling at four sites. Data collected over this period will track all flow events in-stream to assist in validation and calibration of the models into the future.

4 DESCRIPTION OF THE REGION

QUEENSLAND MURRAY DARLING BASIN

The Queensland section of the Murray-Darling Basin covers the Condamine-Balonne, Moonie, Border Rivers, Nebine, Warrego and Paroo Catchments. This covers the areas of the Natural Resource Management (NRM) bodies of Condamine Alliance (CA), Queensland Murray Darling Committee (QMDC) and South West NRM (SWNRM). Although the Bulloo catchment does not drain into the Murray-Darling Basin, it has been included in this modelling exercise due to the Bulloo being included in the QMDB plan area for Queensland Water Resource Plans, and forms part of SWNRM management area (Figure 1).

The Border Rivers model was extended to a defined catchment boundary outlet in the upper part of NSW as opposed to stopping at the Qld border to enable loads to be extracted from the model at a catchment boundary outlet.

The total modelled area drains 336,900 km^2 of Queensland [\(Table 4](#page-14-0) and Figure 1), west and south of the Great Dividing Range. The catchment has highly variable annual rainfall ranging from 1,250mm in the east to less than 500mm in the west. Annual evaporation ranges from 1600mm in the east to 2,800mm in the west.

The Condamine and Maranoa basins drain into the Balonne Basin, hence the three areas were combined into one model. This model covers 26.3% of the total area [\(Table 4\)](#page-14-0). SWNRM model covers over 50% of the total area.

TABLE 4 - OVERVIEW OF CATCHMENTS

FIGURE 1 –THE QUEENSLAND MURRAY DARLING BASIN CATCHMENTS INCLUDING NRM REGIONAL BOUNDARIES

4.1.1 THE CONDAMINE-BALONNE AND MARANOA CATCHMENTS

The Condamine-Balonne River system [\(Figure 1\)](#page-15-0) is a tributary system rising in the steep slopes to the west of the Great Dividing Range. In the upper reaches, the Condamine River is fed by the major tributaries Oakey Creek, that drains the Toowoomba Plateau, Myall Creek, which drains the north east Darling Downs, Charleys Creek to the North of Chinchilla and Dogwood Creek to the west of Chinchilla. The confluence with Dogwood Creek and Condamine River, midway between Condamine and Surat, becomes the Balonne River. The river meanders southwest through Surat and joins the Maranoa River upstream of St George. Downstream of St George, the river becomes a distributary system in an alluvial fan known as the lower Balonne floodplain.

4.1.2 THE MOONIE AND BORDER RIVERS CATCHMENTS

The Border Rivers [\(Figure 1\)](#page-15-0) is a network of perennial streams that rise in the western slopes of the Great Dividing Range on the Granite Belt and New England Tablelands and together form the headwaters of the Darling River. The Macintyre Brook, Severn River (Queensland), Mole River and Beardy River drains the Inglewood district, Granite Belt, Tenterfield and Deep Water districts in the north. The southern section from north of Glenn Innes to Guyra is drained by the Severn River (New South Wales) and Macintyre River. The confluence of the Severn River (Queensland) and the Mole River becomes the Dumaresq River which forms part of the border between Queensland and New South Wales. The Dumaresq River enters the Macintyre River above Goondiwindi and continues to form the border between the two states. The Macintyre River flows generally west before reaching its confluence with the Weir River, west of Goondiwindi. The Weir River headwaters are located in the Dunmore State Forest south west of Cecil Plains. It is fed by a number of tributaries that drain to an area west of Millmerran and Inglewood and north of Goondiwindi. The Weir River generally flows in a south west direction and combines with the Macintyre River, north of Mungindi, where it becomes the Barwon River. During high flow events water can flow from the Weir to the Macintyre River.

The Moonie catchment [\(Figure 1\)](#page-15-0) is bound to the east by the Border Rivers and the north and west by the Condamine Balonne. The Moonie River rises south west of Dalby and south of Tara and flows generally in a south westerly direction. A number of tributaries contribute to its flow with the largest being Teelba Creek, which joins with Bidgel Creek before joining into the Moonie River upstream of Nindigully. The Moonie River flows into the Barwon River near Mogi Mogi in New South Wales.

4.1.3 THE NEBINE, WARREGO, PAROO AND BULLOO CATCHMENTS

The Nebine, Warrego, Paroo and Bulloo catchments [\(Figure 1\)](#page-15-0) are similar through common characteristics of limited rainfall and high evaporation, via streams that cover long distances of flat country, making these catchments an ephemeral set of streams. Nebine, Warrego and Paroo drain into the Darling River in NSW. The Bulloo is a closed system that terminates in the Bulloo Lakes in NSW. The Warrego starts from the south western regions of the Carnarvon Gorge so benefits with a slighter higher level of rainfall off this area. The Langlo and Ward Rivers meet and then intersect with the Warrego merging three streams above Charleville. This flows through to Cunnamulla where it splits into Warrego Minor and Cuttaburra Creek which travels across the border into NSW as two streams. Paroo begins from lowland headwaters and flows over the flat country through to the NSW border. The Nebine catchment is made of several streams that don't connect until after the border.

 4.2 **SOILS AND LANDSCAPES**

The study area is distinct in soils and landscape moving east to west, however, a large amount of common features are shared throughout the broader catchments, especially on the lower flat plains. Soil type is strongly related to geology. The Condamine catchment is heavily influenced by basalts, with a variety of shallow to deep clay soils (Vertosols, Dermosols and Ferrosols) formed in the uplands and alluvia, extending from the headwaters above Warwick to Chinchilla at the lower end of the Condamine floodplain (Figure 2). The Ferrosols are found on the more weathered zones around Toowoomba.

Stoney Kandosols, Dermosols and Sodosols are found on steep terrain of granites and metamorphics in the Granite Belt and Traprock in the south-east of the study area. Unweathered Cretaceous sediments throughout the region yield extensive areas of clay soils (Vertosols and Dermosols) in both uplands and derivative alluvia. Deeply weathered components of the same geologies are covered in shallow stony Kandosols and Tenosols in steeper areas grading into deep Kandosols and Chromosols in flatter areas and in derivative alluvia.

Elements of the Cretaceous geologies and the majority of the eastern Jurassic sedimentary outcrops yield sodic texture contrast soils (Sodosols and Kurosols). Sandy soils (Tenosols, Kandosols) are formed on quartz sandstones and related alluvia in the headwaters of the Maranoa and extending throughout the South Western Catchments.

FIGURE 2 - QMDB SOILS

 4.3 **RAINFALL & HYDROLOGY**

4.3.1 THE CONDAMINE-BALONNE AND MARANOA CATCHMENTS

The Condamine, Balonne and Maranoa catchments have a summer dominant rainfall pattern. The mean annual rainfall decreases from 1,250mm east of Warwick, near the headwaters of the Condamine River, to less than 500mm south west of St George, near the Queensland–New South Wales border [\(Figure 3\)](#page-18-2). There is significant annual variability in rainfall, particularly in the western part of the basin. Summer rainfall is dominated by high intensity storms from October to December, which may be localised.

The majority of the catchments river systems and tributaries are ephemeral, flowing only after significant rainfall events or due to discharges from dams, weirs, sewerage treatment plants and industry. Flow in the catchment could be regarded as highly variable with the mean annual flow at St George being 1.3 million ML with a standard deviation of 616,000 ML. Often, long periods of low or no flow are experienced.

Some perennial flow does occur in the headwater of the Condamine and Border Rivers catchments. Long term records, accessed using DNRM water monitoring flow portal (2016), show that the upper reaches of the

Condamine River at Warwick flow 90% of the time, mid and lower reaches of the Condamine-Balonne flows 75% of the time with the Maranoa catchment flowing 40% of the time. Flow becomes more intermittent as average annual rainfall decreases.

FIGURE 3 - QMDB AVERAGE ANNUAL RAINFALL DEMONSTRATING THE VARIATION THROUGHOUT THE CATCHMENTS

4.3.2 THE MOONIE AND BORDER RIVERS CATCHMENTS

Similar to the Condamine Balonne and Maranoa catchments, the Moonie and Border Rivers catchments have a summer dominant rainfall pattern. The largest rainfall occurs on the eastern NSW area of Border Rivers ranging from above 1,100mm per annum to less than 500mm in the west side of the Moonie [\(Figure 3\)](#page-18-2). The variability and intensity of storms described for Condamine is also applicable in these catchments.

A number of tributaries of these catchments are ephemeral toward the western end of the catchment. Perennial flow is present in some of the eastern catchment with the presence of three dams within this catchment to utilise the water. Flows at the end of system demonstrate the difference due to the rainfall, with Border Rivers at Mungindi (Barwon River) flow 80% of the time, while in Lower Moonie it only flows 50% of the time.

4.3.3 THE SWNRM CATCHMENTS

The catchments of Nebine, Warrego, Paroo and Bulloo have the lowest annual rainfall in this study, with the Warrego catchment the only catchment to receive a small area of rainfall higher than 600mm/year [\(Figure 3\)](#page-18-2). Only the catchments Paroo, Warrego and Nebine flow into the Murray Darling Basin, as the Bulloo Catchment is an endorheic or closed catchment, draining only to the Bulloo Lakes in NSW.

These streams can be ephemeral with periods of no flow, and large flushes after rainfall. Warrego at Cunnamulla flows around 45% of the time, Paroo at Caiwarro flows 55% and Bullo at Autumnvale only 60% of the time. Large periods of flow are less than 1 m3/sec with periods of flow greater than this for all end of systems occurring 26- 31% of the time. Flow becomes more intermittent as average annual rainfall decreases.

4.4 **LAND USE**

4.4.1 THE CONDAMINE-BALONNE AND MARANOA CATCHMENTS

The Condamine-Balonne and Maranoa catchments land use is shown in [Figure 4](#page-19-4) and Table 5. The aggregated landuse categories were based on QLUMP data (QLUMP, 2016) refer Table 19 for groupings. It is dominated by grazing (open and closed) (70% of the area) with areas of irrigated and dryland cropping concentrated along the main river in the upper Condamine and Lower Balonne.

FIGURE 4 – THE LAND USE FOR THE CONDAMINE-BALONNE AND MARANOA CATCHMENTS AND THE MOONIE AND BORDER RIVERS CATCHMENTS DEMONSTRATING THE LARGE COVERAGE OF GRAZING AND CROPPING AREAS

4.4.2 THE MOONIE AND BORDER RIVERS CATCHMENTS

The Moonie and Border Rivers catchments land use is shown in [Figure 4](#page-19-4) based on QLUMP data (QLUMP, 2016) Refer Table 20 for groupings. It is dominated by grazing (open and closed) (60% of the area) with areas of forestry and conservations (20% of area) in the upper Moonie and Border Rivers (Figure 4 and Table 5).

4.4.3 THE SWNRM CATCHMENTS

The SWNRM catchments land use is shown i[n Figure 5](#page-20-0) based on QLUMP data (QLUMP, 2016) Refer Table 21 for groupings. It is largely dominated by grazing occupying greater than 95% of the area. Based on experience in the previous SWNRM model, it was helpful to delineate the grazing areas further. This was done using land types and created the following Grazing – Alluvial, Grazing - Hard Country, Grazing – Sandplains, Grazing - Woodlands/Forests Grazing – Other, all for both open (cleared) and closed (timbered) areas. An overview of the process can be found in section [11.2.3](#page-67-0) (Figure 5 and Table 5).

FIGURE 5 - SWNRM CATCHMENTS LAND USE

Land use across the modelled area is dominated by cattle and sheep grazing in open and forested grazing environments (>70%) (Figure 5). Significant areas of the Border Rivers and Condamine are utilised for dryland and irrigated cropping. An area of 11,901 km² is cropped in the Border Rivers and 10,590 km² in the Condamine. Dryland cropping is the predominant cropping system covering 17% of the total area with irrigated cropping representing only 2.9% of the modelled area.

5 METHODS

The model was built in the Source Catchments modelling framework. Source Catchments is a water quality and quantity modelling framework that has been developed by eWater Ltd. This framework allows users to simulate how catchment and climate variables (such as rainfall, land use, management practice and vegetation) affect runoff and constituents, by integrating a range of models, data and knowledge. Sub-catchments are the basic spatial unit in Source Catchments. A sub-catchment is further delineated into 'Functional Units' (FUs) based on common hydrological response or land use (eWater Ltd 2013). Source Catchments supersedes the E2 and Water CAST modelling frameworks (eWater Ltd 2012) used by previous iterations of water quality models developed for the QMBD area (refer Table 2 - [Previous Model \)](#page-12-1). The major distinction between this model and previous QMDC models is that this model incorporates the dynamic SedNet plugins used in the Paddock to Reef modelling program (Waters et al., 2014). The dynamic SedNet plugin allows the user to represent hillslope, gully and streambank erosion processes as well as in stream and floodplain deposition/re-entrainment (Figure 7).

 5.1 **WATER QUALITY MODEL STRUCTURE**

In the base eWater Source modelling framework, there are two modelling components assigned to each FU representing runoff and Constituent generation. Nodes and links represent the stream network and runoff and constituents are routed from a sub-catchment through the stream network via nodes and links (Figure 6). However, the basic eWater model structure does not represent gully and streambank erosion processes which are important contributions to erosion in the QMDB. Therefore, the approach used in the Great Barrier Reef modelling program (Waters et al., 2014) was adopted. (Refer section 9.4 for full explanation of constituent generation approach).

FIGURE 6 - EXAMPLE OF A FUNCTIONAL UNIT (FU) AND NODE-LINK NETWORK GENERATED IN SOURCE CATCHMENTS. THESE COMPONENTS REPRESENT THE SUB-CATCHMENT AND STREAM NETWORK

5.1.1 CONCEPTUAL APPROACH FOR CONSTITUENT GENERATION

Source Catchments framework allows specific customised models to be added as 'plug-ins' to meet a particular modelling objective. The GBR Source plug-in (Dynamic SedNet) (Ellis and Searle, 2014) was used for the QMBD Water Quality models.

While the GBR Source Catchment models incorporated paddock scale modelled outputs into the model, the QMDB Water Quality model did not use this functionality. Instead, the RUSLE model for cover dependent land uses (For example: grazing, conservation and forestry) using a modified version of the Universal Soil Loss Equation (USLE) Renard et al. 1997, and Event Mean Concentration/Dry Weather Concentration (EMC/DWC) models were used for the remaining land uses. The decision was made to use RUSLE as opposed to paddock model simulations for the QMDC model due to the time constraints, limited capacity to generate paddock model scenarios and there was no requirement to reflect management differences for this project.

In addition, SedNet/ANNEX (Wilkinson et al. 2004) modelling functionality has been incorporated to generate gully and streambank erosion and floodplain deposition, within the daily time-step model (Ellis & Searle 2014, Wilkinson et al. 2014). Thisincluded the daily disaggregation of long-term average annual estimates of gully and streambank erosion. The disaggregation of the long-term load estimates should be treated with caution, given outputs at a sub annual resolution will not necessarily match observed event estimates in the catchments due to the disaggregation approach.

Point source inputs of pollutants from major sewage treatment plants (STP) are included in the model. Losses from the stream as irrigation extractions were also represented at relevant nodes in the model as a daily time-series of flow and concentration. In-stream transport process such as deposition of sediment and particulate nutrients were also represented [\(Figure 7\)](#page-22-0). A more detailed description of the modelling methodology and algorithms are available in Ellis & Searle (2014) and Wilkinson et al. (2014).

FIGURE 7 - CONCEPTUAL DIAGRAM OF DYNAMIC SEDNET SOURCE CATCHMENTS FRAMEWORK

5.1.1.1 MODEL STRUCTURE

Due to the large area modelled, three separate models were created, largely based on the Regional Natural Resource Management (NRM) bodies along with the natural drainage within the area. This simplified the QMBD model into the following individual model builds:

1. South West NRM (SWNRM) model

Four western catchment form the SWNRM model which includes the Nebine, Warrego, Paroo and Bulloo catchments. These catchments are covered by the South West Natural Resource Management group (SWNRM). These four catchments drain into NSW.

2. Condamine Balonne Maranoa (CBM) Model

The rivers of Condamine, Balonne and Maranoa form a single model. Condamine and Maranoa both drain into the Balonne River that flows across the border into NSW. Two NRM bodies cover this area, Condamine Alliance (CA) for the Condamine River (from the Condamine headwaters to the town of Condamine) and the Queensland Murray-Darling Committee (QMDC) from the town of Condamine west, covering the Maranoa and Balonne Rivers.

3. Moonie and Border Rivers (BDMN) Model

The Border River and Moonie model are two catchments that join together below the NSW Border. As the McIntyre River, which is the major river in the Border Catchment, forms the NSW and Queensland border (thus the "Border Rivers Catchment"). This model captures the NSW area of the Border Rivers catchment and ends where the Moonie and Barwon Rivers meet. This catchment area is covered by QMDC NRM group.

The land area covered by each of these models align with the Water Plan and Water Resource Plan areas under State and Federal water planning processes respectively. This is of benefit to the development of HWMPs/WQM Plans as modelling results specific to plan areas will increase the useability of modelling results.

5.1.2 LAND USE FUNCTIONAL UNITS

Functional Units (FUs) for the modelled regions were defined from land use mapping from a number different sources due to the study area extending into New South Wales.

The original detailed land use categories were reclassified into 10 -12 aggregated land uses/FUs to represent the dominant land uses of interest. These were also informed by the previous models FUs [\(Table 2\)](#page-12-1). For this model, some changes were identified to represent the updated land use mapping. This resulted in the land use categories being identical for the Border Rivers and Condamine models. The land use for the SWNRM region was similar with the majority (>80%) being grazing [\(Table 2\)](#page-12-1).

Grazing areas were spilt into open (cleared) and closed (timbered) to enable differences in runoff and constituent generation for the two areas to be reflected in the model. To differentiate between open and closed grazing, closed areas were those areas with a Foliage Protective Cover (FPC) ≥20% (National Committee on Soil and Terrain 2009). Differentiation was made between these two grazing systems to enable representation of different hydrological response units during calibration. Any given land use within a sub-catchment is aggregated and represented as a single area in the model hence is not represented spatially within a sub-catchment. A complete overview of the land use categories can be found in the Appendix [\(11.2\)](#page-63-0).

5.1.3 SUB-CATCHMENT GENERATION

The QMDB Source Catchments models contain nine catchments: SWNRM (the Bulloo, Paroo, Warrego, and Nebine), CBM (Condamine, Balonne, and Maranoa), and BDMN (Moonie and Border Rivers catchments). The catchments were delineated into smaller sub-catchments using a 100 metre, hydrologically enforced Digital Elevation Model (DEM). A stream network, generated from the DEM was used to determine the location of nodes and links. Nodes are located along the network at sub-catchment outlets and include user specified nodes such as stream gauging stations or storages. For each model, a minimum drainage threshold [\(Table 6\)](#page-24-0) was used to identify the major stream network and contributing sub-catchments. This minimum drainage threshold was also based on input from the local NRM bodies should they wish to use this model on completion.

Some further manipulation of boundaries in GIS was done to accommodate the large flat areas between the Balonne, Moonie and Border Rivers catchments. This was done on visual inspection of imagery and local knowledge. The final sub catchments used for the QMDB models can be found in [Figure 8.](#page-25-0)

FIGURE 8 – AN OVERVIEW OF THE QMDB SUBCATCHMENT, WITH ASSOCIATED LINKS AND NODES. THE LINK AND NODES PROVIDE THE REPORTING POINTS THROUGH THE MODEL. THESE ARE ALSO GROUPED AT THE END OF THE CATCHMENT TO REPORT ON TOTAL CATCHMENT EXPORT FOR AREAS THAT HAVE MULTIPLE RIVERS THAT CROSS THE QLD/NSW BORDER

5.1.4 CLIMATE SIMULATION PERIOD

A 36 year climate simulation period was chosen (01/01/1980 – 31/12/2015). This period included a range of extreme wet and dry periods which is an important consideration for hydrology calibration.

Daily climate input files generated for each sub-catchment were used to calculate daily runoff. Rainfall and Potential Evapotranspiration (PET) inputs were derived from the Department of Science Information Technology Innovation (DSITI) Silo Data Drill database (Queensland Government 2014). The data drill accesses grids of data derived by interpolating the Bureau of Meteorology's station records. The data are supplied as a series of individual files of interpolated daily rainfall or PET on a 5 km grid. Source Catchments interrogates each daily grid and produces an 'averaged' area weighted continuous daily time series of rainfall and PET data for each subcatchment.

 5.2 **RUNOFF GENERATION**

The Sacramento rainfall runoff model, one of six available in eWater Source, was used to generate runoff. Storage dynamics (dams/weirs) were simulated as well as irrigation extractions, channel losses and inflows such as sewage treatment plant discharges through specific node models.

5.2.1 HYDROLOGY CALIBRATION PROCESS USING PEST

Hydrology calibration is a major aspect of constituent load modelling given that constituent generation is driven by rainfall and runoff. Thus it was imperative that the hydrology calibration process was rigorous, and achieved the best possible results. For calibration of parameters, the command-line Source model was coupled with a model-independent parameter estimation tool called PEST (Doherty 2005). PEST was set up to use one of its parameter global optimisers, the CMAES (Covariance Matrix Adaptation Evolutionary Strategy) to estimate the optimised value of the 21 hydrological parameters. PEST's CMAES optimises model parameters using automated search algorithms that minimise the difference between modelled and measured flows, i.e. the objective function. In this calibration, we used an objective function introduced by Coron *et al.* (2012). Lerat et al. (2013) further modified this objective function to reduce the volume difference between the simulated and observed total flow volumes and the misalignment of observed and simulated peak flow timing through its three function terms. The modified objective function comprised three terms which aimed to ensure that the total flow difference was within ± 10%, that the high flow peaks were well represented and that the timing and duration of events was also well represented. A fourth term was added to improve the modelled baseflow proportion. The baseflow term minimises the difference between the simulated baseflow and an "observed" baseflow proportion derived from gauged flow data using the Lyne & Hollick (1979) approach which applies a smoothing filter to daily observed flow to split daily flow into baseflow and quickflow. Note final Sacramento parameters used are listed in Tables 16 – 18.

Once calibration was completed, model performance was assessed for the 38 QMDB gauges used in the calibration process. Performance was assessed for the calibration period 01/01/1980 – 31/12/2015. Most gauges had the complete flow records for the entire years through the calibration period. This was applied to the period of data available at each gauging station. This meant that for some sites that had long term data, the calibration period was completed for 36 years, however for some sites this period was 15 years. These sites with shorter calibration periods were in the minority. The model performance was assessed against observed flow data using the criteria i[n](#page-27-3)

[Table 7](#page-27-3) based on Moriasi (2007, 2015) which outlines the evaluation criteria for model performance for monthly flow.

TABLE 7 - MORIASI (2007, 2015) MODEL EVALUATION CRITERIA

R 2 - Statistical measure of how close data fits the 1 to 1 line

NSE – Nash Sutcliffe is similar to coefficient of determination and used to assess predictive power of a model PBIAS – percent bias is a measure of difference between measured and modelled value

RSR – Ratio of root mean square to standard deviation

5.2.2 REGIONALISATION OF CALIBRATION PARAMETER SETS

To further simplify the number of adjustable parameters during calibration, land uses/FUs deemed to have similar hydrologic response characteristics were grouped into two broad 'hydrologic response units' (HRUs); namely 'forested' areas, and cleared or 'non-forested' areas. These broad groupings were selected from previous research in Queensland which suggested these land uses have measurably different drainage and runoff rates given the same climate and soils (Thornton et al. 2007, Yee Yet & Silburn 2003). Flow routing models were also grouped according to the calibration regions. FUs, links and nodes continued to operate as interconnected units within the Source Catchments structure. A calibration region is defined as the region upstream of a gauge or the area between gauges. Each gauging station included in the calibration represented its catchment area, based on the contributing flow to a gauge. Nested gauges (gauged upstream or downstream by other gauges) had contributing areas minus the contributing area of the upstream gauge. The nearest neighbour approach was used to derive parameters for ungauged sub-catchments (Chiew & Siriwardena 2005, Zhang & Chiew 2009). After flow calibration, the parameter sets were applied to each sub-catchment which included the ungauged areas.

5.2.3 STREAM GAUGE DATA SELECTION FOR CALIBRATION

Flow data was extracted from DNRME Hydstra Surface Water Database to provide the 'observed' flow values for calibration. Additional flow data was received from the DSITI Water Quantity modelling team for the Border River catchment as Qld and NSW were working on an updated Integrated Quantity and Quality Model (IQQM) for Border Rivers.

Gauging Stations were identified as suitable for PEST calibration using the following criteria:

- Located on the modelled stream network;
- Had a minimum of 10 years of flow record (post 1980) with suitable corresponding quality codes in the DNRM database;
- An appropriate spatial distribution throughout the catchment; and
- In discussion with DNRME Senior Hydrographer, gauging stations were rated based on their long-term reliability of ratings with a range of events within the data set

Final gauges used in the PEST calibration process are shown i[n Figure 1.](#page-15-0)

 5.3 **CONSTITUENT GENERATION**

5.3.1 WATER QUALITY CONSTITUENTS MODELLED

The water quality constituents modelled were Total Suspended Sediment (TSS) and Total Nutrients(Total Nitrogen and Total Phosphorus). Rainfall and ground cover are two dominant factors affecting hillslope runoff and erosion in the QMDB. Given grazing occupies over 70% of the QMDB, it was important that the models chosen were able to reflect the dominant erosion processes occurring in these landscapes and the spatial variability observed across such a large area. Dynamic SedNet incorporates daily rainfall, spatially and temporally variable cover to generate hillslope erosion. Gully and streambank erosion and floodplain deposition processes have also been represented.

SEDIMENT GENERATION MODELS

5.3.1.1 HILLSLOPE SEDIMENT AND NUTRIENT GENERATION

A modified version of the Universal Soil Loss Equation (USLE) was used to generate hillslope erosion in grazing lands (Renard et al. 1997, Lu et al. 2001, Renard & Ferreira 1993). This modified version is based on the Revised Universal Soil Loss Equation and is referred to as the RUSLE in this document (Lu et al. 2001, Renard & Ferreira 1993). The RUSLE model was chosen due to its proven ability to provide reasonable estimates of hillslope erosion worldwide, including the application in GBR Paddock to Reef models (McCloskey, 2015), the ability to apply the model across a large spatial extent and at the same time incorporate detailed spatial and temporal data layers including cover and rainfall components. For a detailed explanation of the application of the RUSLE model refer to Ellis and Searle (2014).

Hillslope particulate nutrient generation was derived as a function of the clay fraction of the daily RUSLE soil loss, the surface soil nutrients (total nitrogen and phosphorus) concentration and an enrichment ratio (Young, Prosser and Hughes, 2001). This algorithm assumes that all particulate nutrientsin the soil are attached to the clay faction where:

EQUATION 1

Hillslope particulate nutrient load (kg/ha) = RUSLE sediment load (kg/day) x clay (prop) x Surface nutrient concentration (kg/kg) x Enrichment factor x Nutrient Delivery Ration (NDR)

This estimates the total suspended nutrient load which reaches the stream. The dissolved nutrient load is the product of an EMC/DWC value and the quick and slow flow respectively. These models are described in (Ellis and Searle 2014) and replicate the original SedNet approach to dissolved and particulate nutrient generation modified to a daily basis.

Gully generation model

To derive gully erosion estimates, the critical input data layer in the gully model is gully density. This was generated through a methodology created within DNRM by which a square kilometre is divided into one hundred 100 square metre plots. These plots were examined for the presence of a gully. Where a gully was present that plot was marked as a positive for gully. The square kilometre was scored based on the number of gully plots as a percentage. Gully mapping was conducted as part of this project for 6,500 square kilometres the QMDB on a relatively even grid to sample all land types within catchments. Using cover and land types, the average gully density was extrapolated for the entire QMBD. The extrapolation provided an average gully density for that land type and cover amount, which provided a complete gully density map for the QMDB. Further information on this process can be found in Darr (2017).

5.3.2 OTHER LAND USES: EVENT MEAN CONCENTRATION, DRY WEATHER CONCENTRATION

The remaining land uses: forestry, nature conservation, urban, horticulture and the 'other' land use category had EMC/DWC models applied. In comparison to grazing and cropping, these land uses had a small relative contribution to regional loads. In the absence of specific models for these land uses, EMC/DWC models were applied where daily load is:

EQUATION 2

Daily Load (kg) = (EMC (mg/L) x quickflow runoff (ML)) + (DWC (mg/L) x slowflow runoff (ML))

Where quickflow represents the storm runoff component of daily runoff, the remainder was attributed to slowflow. EMC/DWC values were derived from monitoring data, or where monitoring data was not available, from previous studies (Waters & Packett 2007, Rohde et al. 2008, Bartley et al. 2012, and Turner et al. 2012) - Refer to [Table 23](#page-73-1) in the appendix for a summary of EMC/DWC values used.

The EMC/DWC values were based on median values used in the Waters, 2008 report for another SWNRM source model. The only change from these figures used was to increase the phosphorus median values for the BDMN model by 3 times to better represent the values that was being observed and predicted.

These median values were then spilt into ratios for Nitrogen and Phosphorus species for the hillslope model. This was based on the laboratory samples taken throughout the QMDB continuous monitoring project (section [5.4.1.4\)](#page-36-0) which assigned an average proportion for DIN, DON and particulate N, FRP, DOP and Particulate P from the known laboratory sample proportions.

5.3.3 NODE BASED MODELS

Nodes represent points in a stream network where links are joined or catchment outlets (eWater Ltd 2013). Catchment processes can be represented at nodes. For a detailed description of how these models work refer to the Source Catchments Scientific Reference Guide (eWater Ltd. 2013). In the QMDB Catchments models, irrigation extractions, sewage treatment plant (STP) inflows and storages/weirs were represented at nodes. The following sections provide a brief outline of how these models were applied.

5.3.3.1 EXTRACTION, INFLOWS AND LOSS NODE MODELS

To simulate the removal of water and the associated load of constituents from storages and or rivers, daily extraction estimates for a river reach were incorporated at relevant nodes. The irrigation extraction data was obtained from Integrated Quantity and Quality Model (IQQM) runs provided by Queensland Hydrology (DSITI) for each region. Multiple types of extractions were aggregated and allocated at the appropriate downstream nodes. Regionally specific loss models were included to account for channel losses where necessary. To account for water moving between channels at bifurcations in the rivers (e.g. the split of the Condamine River into the North and South Branch, Macintyre River in the lower Border Rivers and the lower end of the Warrego River in Southwest NRM region). Loss nodes and inflow nodes were added to the model to mimic discharge out of the main river (loss node) and this water entering a bifurcation or branch downstream (inflow node).

5.3.3.2 STORAGES

Storages (dams and weirs) with a capacity >10,000 ML were incorporated into the model at nodes. Only storages of significant capacity were incorporated as it was impractical to include all storages and it was assumed the smaller storages would have minimal impact on the overall water balance and pollutant transport dynamics. Storage locations, dimensions and flow statistics were used to simulate the storage dynamics on a daily basis. Trapping of fine sediment and particulate nutrients were simulated. Fine sediment and particulate nutrients were captured using a 'trapping' algorithm based on daily storage capacity, length and discharge rate (Lewis et al. 2013). Dissolved constituents were decayed in storages using a first order decay model.

5.3.4 IN-STREAM MODELS

The in-stream process models can represent streambank erosion, in-stream deposition, decay and remobilisation of fine and course sediment and particulate nutrients and floodplain deposition. The following sections provide a brief outline of their application.

5.3.4.1 SEWAGE TREATMENT PLANT (STP) INFLOWS

Those centres that had an STP that was 10,000 equivalent persons or higher had the average annual nutrient load discharged from the STP added into the source model at the relevant link. For the QMDB, the only STP that met this criteria was Toowoomba's Wetella sewage treatment plant. Data was accessed from the Department of Environment and Science STP database and the average annual discharge prior to and post the upgrade were used.

5.3.4.2 STREAMBANK EROSION

The streambank erosion model implemented is based on the SedNet modelling approach (Wilkinson et al. 2010). A mean annual rate of fine streambank erosion $(t/\gamma r)$ is calculated as a function of riparian vegetation extent, streambank erodibility and retreat rate. The mean annual streambank erosion was then disaggregated as a function of the daily flow. For a full description of the method refer to Ellis & Searle (2014).

For particulate nutrients, particulate N and P contribution from streambanks was estimated by taking the mean annual rate of streambank erosion (t/yr) multiplied by the Australian Soil Resource Information System (ASRIS) subsurface soil N and P concentrations. The mean annual streambank erosion was then disaggregated as a function of the daily flow.

5.3.4.3 FLOODPLAIN DEPOSITION

When floodwaters rise above river banks the water that spills out onto the rivers floodplain is defined as overbank flow. Floodplain trapping or deposition occurs during overbank flows. The velocity of the flow on the floodplain is significantly less than that in the channel allowing fine sediment to deposit on the floodplain. The amount of fine sediment deposited on the floodplain is regulated by the floodplain area, the amount of fine sediment supplied, the residence time of water on the floodplain and the settling velocity of the sediment (Prosser et al. 2001, Wilkinson et al. 2010, and Ellis & Searle 2014). For particulate nutrients, the particulate nutrient load deposited on the floodplain is a proportion of fine sediment deposition. The loss of dissolved nutrients on the floodplain were not modelled. Details on the floodplain deposition and remobilisation models can be found in Ellis & Searle (2014).

5.4 **ASSESSMENT OF CONSTITUENT LOAD PERFORMANCE**

For load calibration, monitored TSS, TN and TP samples were correlated against discharge for historical water quality data from Qld and NSW datasets to provide a long-term comparison (30+ years) of catchment loads[\(Figure](#page-15-0) [1\)](#page-15-0).

For independent validation, a short-term comparison (2015-2016) between observed and modelled concentrations was made using data collected from a project funded concurrently with this project called DEHP5 monitoring Project where water quality data was collected at four gauging stations established to support this project.

5.4.1 LOG-TRANSFORMED WATER QUALITY RELATIONSHIP ESTIMATOR (1980-2015)

Monthly and annual sediment and nutrient load estimates were derived from monitoring data to calibrate the QMDB Source Catchments model for the period January 1980 to December 2015 (36 years). Historically, within the QMDB catchments water quality data was collected sporadically and often not sampled for critical parts of the hydrograph. Efforts have been made to capture the range of values over the hydrograph through novel approaches using Water Quality sampling trailers (DNRME/SWNRM) and targeted event sampling between 2003 and 2008, dramatically increasing the event samples collected in SWNRM region with an additional 50 samples collected, which counts for 70% of the event samples for this catchment (Waters, 2008). Water quality samples were accessed from DNRME's Hydstra data set. A summary of the data is shown in [Table](#page-31-2) 8.

TABLE 8 – SUMMARY OF WATER QUALITY DATA FROM DNRM'S HYDSTRA DATABASE. THE RESULTS HIGHLIGHT THE VARIABILITY IN THE NUMBER OF SAMPLES CAOLLECTED IN EACH CATCHMENT AND THE DIFFERENCES IN CONCENTRATIONS BETWEEN EVENT AND LOW FLOW CONDITIONS.

To calibrate the Source models, ambient and event water quality data was log transformed and plotted against log transformed flow. Curves were then fitted to derive relationships between TSS, TP and TN concentrations and gauged discharge by catchment. The log transformation was undertaken to enable the data to become more symmetric and normally distributed and to make the data variance more uniform. This approach was based on an analysis of the original data and finding a positive skew and large difference in the data variance. It has become common practice to log-transform water quality data, especially chemical concentrations and stream discharge, because this simple transformation often fits the inherent assumptions when using regression analysis (Li and Migliacio, 2010). Seasonal fluctuations in concentration and short-term fluctuations related to fluctuations in flow are two factors that greatly increase variance and hinder trend detection (Richards and Baker, 2002). Removing time dependent factors allows for the relationship between flow and concentration outside of these fluctuations.

This relationship was applied to the hourly flow time series. Concentrations could then be calculated at a range of gauging stations across the region where hourly flow data was available. The resulting concentration values were used to calculate an hourly load and aggregated to provide an estimate of monthly and yearly load at validation sites throughout the catchments. Due to limited availability of measured reference data at a full range of flow heights in the data used to build the relationship, it should be noted that there is a degree of uncertainty about the measured load estimates at high flows. However, in a data poor catchment this provides the most robust approach to calibrate and validate the models. Water quality and flow data were vetted to remove erroneous data points. The overall soundness of the estimates were cross checked by calculating an average annual load by as product of the average constituent concentration (pooling all data) and the average annual flow at a site. The resultant figure gave a logic test to ensure the relationship created wasn't providing erroneous results.

5.4.1.1 LOG-TRANSFORMED TSS VS DISCHARGE RELATIONSHIP

Separate curves were derived for the three catchments correlating TSS to discharge (see [Figure 9,](#page-33-0) [Figure 10](#page-33-1) and [Figure 11\)](#page-34-0). The data available for a catchment informed whether a transformation using log or log +1 was undertaken and then these were assessed on the regression relationship that was achieved. For the Condamine Balonne model and by the process of testing the soundness of the result, it was revealed that in higher flows, the linear relationship was providing too much TSS based on a logic test for expected TSS loads. This logic test used an average concentration from the data applied to the flow, which provided a normal range. The linear relationship was successful at low and medium flows, however overestimated the TSS loads at large flow/events. This was rectified through using a polynomial relationship that limited TSS concentrations at high flows.

The strength of the relationship for each models data is found in [Table 9,](#page-32-0) which demonstrates while the R Squared value may be low, the relationship was significant. The number of observations come from how successfully a flow value was able to be assigned to a constituent. It can be seen from Table 12 that there was a large number of observations in the Condamine Balonne Maranoa catchments in contrast to the South West and Border Rivers-Moonie catchments.

TABLE 9 - REGRESSION TESTS SUMMARY FOR DISCHARGE RELATIONSHIPS

FIGURE 9 – LOG-LOG TSS DISCHARGE SWNRM RELATIONSHIP

FIGURE 10 - LOG+1 - LOG+1 TSS DISCHARGE MNBD RELATIONSHIP

FIGURE 11 - LOG+1 - LOG+1 TSS DISCHARGE CBM RELATIONSHIP

5.4.1.2 LOG-TRANSFORMED TN AND TP VS DISCHARGE RELATIONSHIPS

To derive load estimates for nutrients, relationships were derived between discharge and TN and discharge and TP for each of the regions following the same methodology as used to derive discharge and TSS relationships.

Given the strong relationship between TSS vs TN and TSS vs TP found by Waters et al. (2008), due to the high particulate fraction in runoff for these basins, an additional correlation was derived between TSS vs TN and TSS vs TP to give another estimation of the nutrient load. This was included as only TP and Discharge had a higher number of observations than other counterparts [\(Table 10\)](#page-34-1) and weaker relationships (R^2 and significance) with discharge [\(Table 10\)](#page-34-1). Using two sets of measured data provides a potential range for the TN and TP values, ultimately trying to address the lack of available data by building a reliable load estimate from observed data to calibrate against. Section [11.8](#page-92-0) contains all plots of the correlations described above.

REGION		Relationship Base	Equation	R Square	Observations	Significance F
SWNRM	ΤN VS	Discharge	$v = 0.102x + 0.0858$	0.27740	83	3.14643E-07
		TSS	$y = 0.3448x - 06911$	0.67058	101	1.30646E-25
	ТP VS	Discharge	$y = 0.1676x - 0.4993$	0.45868	109	6.17562E-16
		TSS	$y = 0.4981x - 1.5967$	0.74451	152	2.6842E-46

TABLE 10 - REGRESSION TEST SUMMARY FOR TN/TP RELATIONSHIPS PROVIDING THE RELATIONSHIP EQUATION USED FOR THE CALCULATION OF AN ESTIMATE FROM 'OBSERVED' DATA FOR LOAD CALIBRATION SITES

5.4.1.3 LONG-TERM FLOW RANGE CONCENTRATION ESTIMATOR (1980–2015)

Sediment and nutrient loads were estimated at both the annual and monthly time steps from the relationships described above i[n Table 10.](#page-34-1) These estimates are referred to as the "observed" loads from this point on, and refers to an estimate derived from observed sample data used to validate the QMDB Source Catchments model for the period January 1980 to December 2015 (36 years).

A subset of calibration sites used for the Hydrological Calibration [\(5.2.3\)](#page-27-1) were used for the TSS, TN and TP calibration process. These sites were selected because they were located at the end of major basins (for example Condamine, Balonne Warrego) (Figure 1) to capture broad representative areas of catchment land use and features that were assumed to have a similar response to the TSS, TN and TP generation. In addition there was little extra monitoring data available at other sites to justify calibrating at a finer scale in any basin.

Calculation of monthly loads enabled a consistent statistical model evaluation for sediment and nutrients using Moriasi et al. (2007, 2015). Four quantitative statistics were used [\(Table 11\)](#page-35-0). The statistics were calculated and model performance rated.

An important element of the calibration was to adjust broad spatial parameters of regions that generated the load for TSS. This included streambank erosion parameters (height of bank, width, and overbank flow reoccurrence interval), hillslope fine sediment delivery ratio and for gully erosion, the average gully cross section and the delivery ratio of these gullies. Gully width and streambank height were derived from aerial photography at approximately 50 sites across the region. This number of sites were sampled as it was deemed to provide a reasonable spatial coverage for the limited time and resources available to complete the project.
5.4.1.4 INDEPENDENT DATA COMPARISON USING GRAB SAMPLE DATA

In addition to the long-term load comparisons, a short-term comparison (2015-2016) of modelled and observed constituent concentrations was made using grab sample data collected from the previously mentioned monitoring project running concurrently (DEHP5). The project uses a combination of grab samples and continuous monitoring probes at four gauging stations throughout the QMDB catchments to test whether a continuous monitoring system could reliably provide TSS and EC values with the long term aim to derive correlations between turbidity and TSS/TP/TN and EC and total dissolved solids for continuous load estimates. This project is being run over three years (2016-2018) and involves regular site visits to calibrate these probes. During these visits and during runoff events grab samples are taken to correlate to the probe readings.

6 RESULTS

6.1 **HYDROLOGY CALIBRATION RESULT**

The Moriasi et al (2007, 2015) results for the summary of all calibration gauges for the models are found in [Table](#page-38-0) [12](#page-38-0) with full detailed results in Table 22. Condamine Balonne Model returned the best result using the criteria set out in Table 11, scoring "Very Good" for 100% of measures for all of its catchments. This was followed by SWRNM which had 85% of all measures being "Very good" and Border Rivers with 78% of all measures being "Very good".

Overall, PBIAS was "Very Good" for 97% or 36 of the 37 calibration gauges for the volume of modelled versus the observed over the modelled period. 81% of gauges scored a "Very Good" for R2 monthly flow. The monthly NSE was "Very Good" for 78% of the gauges and RSR had "Very Good" for 89% of the gauges calibrated.

Scatter plots showing predicted and observed monthly stream flow are presented in [Figure 12](#page-39-0) - [Figure](#page-39-1) **15**. These plots demonstrate the relationship of observed and predicted flow, and the correlation of the predicted monthly flows to the monthly observed flows.

[Figure 16](#page-40-0) shows a typical cross verification of daily observed runoff (blue line) and predicted runoff (red line). These are from a range of events and gauging stations. The hydrographs demonstrate the good alignment between predicted and observed flow.

TABLE 12 - SUMMARY OF MONTHLY HYDROLOGY CALIBRATION AGAINST MORIASI (2007, 2015) PERFORMACE EVALUATION CRITERIA

the remaining Paroo calibration site results: R2 scored "Good", NSE and RSR scored "Satisfactory" for Moriasi Performance Criteria

^ The remaining Warrego calibration site results: R2 and RSR scored "Good", NSE scored "Satisfactory" for Moriasi Performance Criteria

* The remaining Moonie calibration site results: R2 scored "Good", NSE scored 66% "Good" and 33% "Satisfactory" and RSR scored "Satisfactory" for Moriasi Performance Criteria

@ The remaining Border River calibration site results: R2 scored 22% "Good" and 11% "Satisfactory", NSE scored 22% "Good" and 11% "Satisfactory", PBIAS scored "Satisfactory" and RSR scored "Satisfactory" for Moriasi Performance Criteria

FIGURE 12 – WARREGO, PAROO, NEBINE AND BULLOO MONTHLY CALIBRATION RESULTS

FIGURE 13 – MARANOA AND BALONNE MONTHLY CALIBRATION RESULTS

Figure 14 – CONDAMINE MONTHLY CALIBRATION RESULTS

FIGURE 15 **–** Moonie and Border Rivers MONTHLY CALIBRATION RESULTS

FIGURE 16 - TYPICAL CROSS VERIFICATION OF HYDROGRAPH PEAK ALIGNMENT

6.2 **LOAD CALIBRATION AND VALIDATION**

Load calibration was undertaken through an iterative process of parameterisation of the streambank, gully and hillslope erosion components of the Source models.

Average annual and annual loads were first calculated as per equations derived in section 5.4.1. The modelled loads were then compared to the loads estimated from the observed data. Hillslope delivery ratios or gully cross sectional areas or in stream deposition rates were then adjusted up or down, the model then rerun and the process continued until modelled loads were within a sensible range for the Moriasi statistics to better align with the observed load estimates until the modelled and observed loads were deemed to be suitable from average annual and annual perspective.

The Moriasi results for TSS, TN and TP at the calibration gauges are summarised in [Table 13](#page-42-0) for the yearly comparison, with the monthly summary of sites shown in [Table 27](#page-85-0) calibration results. The yearly calibration performs equal to, or better than, the monthly calibration results in all cases.

Overall, 10% of the sites achieved a TSS calibration score of Very Good and ~65% of performance measures scored above Satisfactory for Total Suspended Solids for Moriasi results. For Total Nitrogen, 56% scored Very Good and ~43% of performance measures scored above Satisfactory for Moriasi results. For Total Phosphorus, 42% of the statistics rated as Very Good and ~57% of performance measures scored above Satisfactory for Moriasi results. Scatter plots showing predicted and observed yearly loads are presented for TSS, TN and TP in [Figure 17-](#page-43-0) 34. Full Moriasi results for both monthly and yearly analysis are provided in Tables 28-29 for TSS, Tables 30-31 for TN and Tables 32-33 for TP.

TABLE 13 – MORIASI ANNUAL LOAD CALIBRATION SUMMARY

FIGURE 17 - SWNRM YEARLY PREDICTED VS OBSERVED TSS LOAD

FIGURE 19 - MARANOA BALONNE YEARLY PREDICTED VS OBSERVED TSS LOAD

FIGURE 18 - CONDAMINE YEARLY PREDICTED VS OBSERVED TSS LOAD

FIGURE 20 –BORDER RIVERS & MOONIE YEARLY PREDICTED VS OBSERVED TSS LOAD

Note: "*Observed" loads were derived from correlations between flow and water quality samples for TSS/TN/TP*

For SWNRM, MB and Condamine the scatter plots shown in [Figure 17](#page-43-0) - [Figure 20](#page-43-1) show a good fit between predicted and observed TSS loads with R^2 all above 0.60. Similarly for TN (Figure 21-24), R^2 are greater than or equal to above 0.87. In general the highest 6-10 events are under predicted by the model and small events are over predicted to compensate for this.

FIGURE 22 - CONDAMINE YEARLY PREDICTED VS OBSERVED TN LOAD

FIGURE 23 - MARANOA BALONNE YEARLY PREDICTED VS OBSERVED TN LOAD

FIGURE 24 – MOONIE AND BORDER RIVERS YEARLY PREDICTED VS OBSERVED TN LOAD

FIGURE 25 - SWNRM YEARLY PREDICTED VS OBSERVED TP LOAD

FIGURE 26 - CONDAMINE YEARLY PREDICTED VS OBSERVED TP LOAD

FIGURE 27 - MARANOA BALONNE YEARLY PREDICTED VS OBSERVED TP LOAD

FIGURE 28 - MOONIE AND BORDER RIVERS YEARLY PREDICTED VS OBSERVED TP LOAD

The scatter plots for predicted vs observed yearly TP [\(Figure 25,](#page-45-0) [Figure 26,](#page-45-1) [Figure 27](#page-45-2) and [Figure 28\)](#page-45-3) show a good fit with the R^2 values all above 0.78, however these fits were not aligned well to the one to one relationship except for SWNRM. All larger loads were under predicted, however for Maranoa Balonne, Condamine and Moonie-Border Rivers this under prediction also began at the mid to large loads, over predicting the smaller loads also in the Condamine and Moonie-Border Rivers. Given loads are a function of flow, a number of the significant large runoff events were under predicted and this was translated through into loads. The exception being TN which suggests that the input data used to generate Total Nitrogen loads may be an over estimation of the actual generation rates.

6.2.1 AVERAGE ANNUAL AND TOTAL LOAD COMPARISON

Results for the average annual loads for TSS, TN and TP were calculated for predicted and observed [\(Figure](#page-46-0) [29,](#page-46-0) [Figure 30](#page-47-0) an[d Figure 31\)](#page-48-0) respectively. Annual loads for all constituents are also provided in Tables 24-26.

6.2.1.1 TOTAL SUSPEND SOLIDS RESULTS

[Figure 29](#page-46-0) shows the average annual predicted and observed TSS loads for each calibration gauging station for the three modelled areas. This figure demonstrates the range of total loads between the three regions, highlighting that Border Rivers/Moonie has a much a lower average annual sediment load (approx. 10% of the SWNRM and CBM models). Important to note is the small contributions for Moonie River at Flinton (417205A) and the Weir River at Talwood (416202A). The load is on average, over predicted by 45% for the Border Moonie (BDMN) Catchments models. For the CBM, the load increases as you move downstream from the uplands at Warwick (423210C) down the Condamine to Cotswold (422325A) and then Weribone (422213A) on the Balonne River. The lower loads generated from the Maranoa basin at the gauge (422404A) are evident. The Maranoa and Weribone gauges both drain into Beardmore Dam. The Culgoa gauge (422204A), is downstream of Beardmore Dam which accounts for the drop in load at Culgoa gauge. In the Condamine basin, the model over predicts the observed loads by an average of 23% with a maximum of 45% at the top gauge (423210C) down to 8% at the furthest downstream gauge in the Condamine (422325A) . For SWNRM, the model over predicts the load with the results from all three gauges within 20% of the observed load.

FIGURE 29 – AVERAGE ANNUAL PREDICTED VS OBSERVED TSS LOADS FOR CALIBRATION GAUGES

6.2.1.2 TOTAL NITROGEN RESULTS

[Figure 30](#page-47-0) summarises the average annual predicted versus observed TN loads for each calibration site. There are two observed estimates of TN, one being correlated against flow and the other against TSS concentration.

For the Condamine Balonne region, modelled TN loads were within 10% of observed for the middle section of the Condamine (422333A – Loudon Bridge) sites, and lower Balonne (422204A) and Maranoa (422404A) sites, while the model is 35% under predicting in the mid-section gauges at the end of the Condamine(422313A) and the top of the Balonne Rivers (422404A). SWNRM modelled loads were within 20% of the observed loads. For the BDMN, the model under predicted (within 30%) the TN load for three of the four load calibration gauges.

FIGURE 30 - AVERAGE ANNUAL PREDICTED VS OBSERVED TN LOAD AT EACH CALIBRATION GAUGE

6.2.1.3 TOTAL PHOSPHORUS RESULTS

Figure 31 shows the average annual observed and predicted TP load at the calibration gauges. The pattern is very similar to TN which is not unexpected given the observed loads are correlated to the same data as TN and both are well correlated to TSS. TP was over predicted for all sites (up to 23%) in SWNRM region. For BDMN, the predicted loads range from 41% over prediction to minimum of 2% under prediction. For CBM, TP loads were over predicted. Loads at the top of the Condamine (423210C-Warwick and 422333A – Loudon) and the Lower Balonne (422204A – Culgoa) and Maranoa (422404A – Cashmere) sites. The lower Condamine (422325A – Cotswold) occurs at the lower end the observed range, and Upper Balonne (422213A – Weribone) under predicts. Overall the difference for CBM was around 50%.

FIGURE 31 - AVERAGE ANNUAL PREDICTED VS OBSERVED TP LOAD AT EACH CALIBRATION GAUGE

6.2.2 CATCHMENT LOADS RESULTS

[Table 14](#page-49-0) shows the constituents exported as a total load and on a per hectare basis. For CBM, Balonne has the largest export of TN, TP and TSS and both Condamine and Maranoa loads are significantly less. This can be explained by the fact that the Condamine, and Maranoa drain into the Balonne catchment above Beardmore dam (Figure 1), thus the load exported from the outlet of the Maranoa catchment includes runoff and loads from the Condamine and Maranoa system.

For the BDMN model, the Border Rivers catchment has a higher export for TSS, TP and TN than the Moonie catchments. For the SWNRM model the catchment exports loads from the lowest for TSS, TP and TN, to the highest respectively are Nebine, Warrego, Paroo and Bulloo. It is interesting to note that for BDMN the TN exported load is around 10 times of that TP, whereas for CBM and SWNRM the ratio of TN export is only around three times greater than the TP export.

TABLE 14 - EXPORT CONSTITUENTS LOADS BY CATCHMENT

6.2.3 SEDIMENT BUDGET FOR TSS

The model is able to track sediment and nutrient sources and sinks to create a sediment budget. [Table 15](#page-49-1) summarises the components of the sediment budget as a proportion of the total load exported for each major basin in the QMDB. Hillslope erosion has the greatest percentage contribution of sediment in the Border Rivers catchment and the lowest in the Balonne catchment. For gully contribution, the highest is Moonie and lowest is the Condamine, and for Streambank, the highest relative contribution is from the Condamine and lowest from Moonie catchments.

TABLE 15 – TSS SEDIMENT BUDGET AS A PROPORTION (%) OF TOTAL EXPORT

6.2.4 LAND USE LOADS

[Figure 32,](#page-50-0) [Figure 33,](#page-50-1) [Figure](#page-50-2) 34 show the breakdown of the contribution to total export by landuse.

The contribution from grazing is the largest contribution to export for TSS for BDMN and SWNRM catchments (~40-80% respectively) while cropping is the greatest percentage of TSS load exported for the Condamine Balonne and Maranoa catchments (~20%). The per unit area contribution by landuse is also provided in Table 34.

The contribution from grazing is the largest for TN and TP SWNRM catchments (>80%) while cropping is the greatest percentage of TN and TP loads exported for the Condamine Balonne and Maranoa catchments and Moonie Border River catchments (>20% in both).

FIGURE 32 – TSS CONTRIBUTION TO EXPORT BY LANDUSE

FIGURE 33 – TN CONTRIBUTION TO EXPORT BY LANDUSE

FIGURE 34 – TP CONTRIBUTION TO EXPORT BY LANDUSE

6.2.5 INDEPENDENT VALIDATION

[Figure 36](#page-52-0) and [Figure 35](#page-51-0) show the modelled and independent grab sample concentrations plotted for 2015- 2016 at Mungindi (416001). These preliminary results are very encouraging and suggests that the predicted TSS concentrations for the 2016 event are showing good agreement with the modelled TSS concentrations particularly at higher TSS concentrations greater than 200 mg/l.

FIGURE 35 – COMPARISON OF LABORATORY SAMPLES AGAINST MODELLED TSS CONCENTRATION AT MUNGINDI FOR THE PERIOD JULY 2015 TO JANUARY 2017

Grab samples of TP and TN are generally half the modelled concentrations [\(Figure 36\)](#page-52-0). Both TN and TP again follow the similar trend over an event and the model follows the drop (01/01/16) on the falling stage of the hydrograph during the event. The modelled and observed concentrations are variable although tend to suggest that during event flows the model is over predicting concentrations. However the results are encouraging given the model is within an order of magnitude of the measured concentrations. Further work is now required to investigate why the model is over predicting. The result could be due to overestimation of the hillslope generation loads or potentially the input layer for TN and TP in the gully and streambank models may have been derived from limited field data. Future work needs to investigate the results further.

FIGURE 36 – INDEPENDENT TN AND TP LABORATORY SAMPLES AGAINST MODELLED TN AND TP CONCENTRACTION AT MUNGINDI

7 DISCUSSION

HYDROLOGY

The hydrology calibration produced a high quality calibration when assessed against Moriasi (2007, 2015) performance criteria with 36 of the 37 gauges calibrated within ±5% of observed flow. The hydrology calibration was a significant improvement on the models previously created by Waters (2008). In the EMSS QMBD model, predicted flows were ±33% using SIMHYD and only 7 of the 14 gauges calibrated met both the required performance measures (Waters, 2006). The monthly NSE values were generally rated as very good with average monthly NSE values averaging 0.87, 0.81 and 0.84 for CB, SWNRM and BDMN respectively.

Where poorer calibrations were achieved it was generally due to the large distances between gauges. For example, in the SWNRM in Warrego where the Ward River joins the Warrego below Charleville there were calibration issues. This was also present in the CBM at the Weribone gauging station where there is a 300km stretch of river between calibration sites with multiple tributaries joining. Even though the calibration process utilised the latest techniques to best fit the hydrograph timing this section was problematic. In the Border Rivers, the issues were more associated with headwater gauges, especially capturing the flows off the areas of granite country. One potential source of uncertainty is due to the highly variable rainfall along the elevated ridges of the Border Rivers where there may be a lack of rainfall gauges to adequately capture the variability.

The final issue was at Mungindi gauging station in the Border Rivers where the Macintyre River joins the Weir River and through the braided section of river after Goondiwindi. This may be attributed to the highly variable flow path of runoff depending on the discharge with multiple break out areas at difference flow stages.

Representing pollutant movement through the braided stream networks such as the low lying areas of the Warrego, Balonne, Border Rivers and the Condamine catchments, created some challenges in the current Source configuration. Losses from one section of the river had to be reflected as an inflow in adjacent drainage lines where it was known that a break out occurred into an adjacent stream above a certain discharge. This approach was necessary at a number of locations to maintain a mass balance. In future modelling, a new plug in could assist or an alternative approach such as aggregation of these areas into a single catchment may result in an easier model build and less complexity in constituent load calculation.

7.2 **MODELLED CONSTITUENT LOADS AND VALIDATION**

7.2.1 CONSTITUENT LOAD CALIBRATION AND VALIDATION

Having limited water quality data to calibrate a model for such a large area was challenging although not an uncommon problem in the water quality area. Calibration of the model therefore required a relationship to be derived between the available sample data and flow which was then used to build a long-term average annual load estimate.

The loads derived from historical measured data provided a useful reference dataset to calibrate against. However, a low number of samples for high flow events introduces possible errors in the relationship. Collecting further samples over the range of the hydrograph may result in greater confidence in the measured load estimates. Overall, the Log-Log relationships provided a satisfactory method to calculate an "observed" set of constituent loads. This was used to enable the comparison of the predicted loads and provided a satisfactory calibration in areas where there is limited reference data to calibrate loads against. This however does not preclude a complete set of reference water quality data to calibrate against when one is available. Comparing modelled concentrations to the independent samples (Figure 35 and [Figure 35\)](#page-51-0), was very encouraging and suggests that the model is doing a reasonable job of load prediction as a first approximation.

When deriving the TSS concentration and flow relationship there were almost double the number of samples available compared to the number of TN and TP samples to derive the correlations(CBM TSS had ~1800 samples, TP had ~1000 samples) (also refer [Table 9](#page-32-0) and [Table 10\)](#page-34-0). The greater number of samples available to derive the TSS vs flow correlation gave greater confidence in the estimate for TSS compared to TN and TP estimates.

The results identified larger discrepancies between predicted and observed loads for event years. Predicted results were in line with the observed loads for average rainfall years. In the absence of TSS and nutrient data at very high flows ideas on how to "cap" the load were tested including setting an upper limit on the TSS concentration. The rationale being that sediment concentrations can often reach a maximum as flow increases which can be attributed to sediment exhaustion from the catchment. To reflect this, a polynomial relationship was used which allowed the concentration of the TSS to flatten off at higher flows [\(Figure 11\)](#page-34-1).

In previous work in SWNRM catchments (Waters, 2008) it was noted that TSS was well correlated to both TN and TP. This relationship was investigated further for datasets outside of SWNRM and the method was repeated for all regions [\(Figure 30](#page-47-0) and [Figure 31](#page-48-0) respectively) to provide an alternative estimate of loads at gauges. The loads derived from TSS/TN and flow/TN were similar. For TSS/TP however the result was markedly different to the flowTP in the CBM with FlowTP being ~20% higher than the TSS/TP relationship. This result occurred in the section of river at Chinchilla gauge (Figure 1). In this area there is extensive cropping suggesting that fertiliser inputs may have influenced the correlation between flow and constituents potentially. Further down the Balonne system where grazing dominates the load estimates using TSS and flow and TSS vs TN and TSS vs TP were much more closely aligned.

The modelled hillslope/gully/streambank erosion proportions seem credible when compared to modelled contributions in the Fitzroy Basin (McCloskey, 2017). The Fitzroy was used for comparison due to its relative similar range of land uses, climate and size as the QMDB catchment. The Fitzroy's sediment loads were 29% from hillslope, 52% from gully and 19% from streambank erosion, with the QMBD relative contributions being 18% from hillslope, 42% from gully, 39% from streambank.

One observation from the modelled load estimates is the contrasting loads exported for each of the basins. The difference between generation rates of sediment between relatively similar areas of CBM (3.5 kg/ha) and BDMN (0.3 kg/ha) with both having similar climatic conditions and land uses may be explained by the differences in runoff between the two regions and secondly, the low TSS concentrations of samples collected in the BDMN. The runoff from the upland areas of the Border Rivers make up around 40% of the BDMN catchment, while the majority of the CBM is flat with close connectivity to the river from cropping lands. The flow from BDMN (600,000 ML/yr) exhibits lower sediment concentrations than the CBM due to its more natural environment of forestry and nature conservation in the uplands areas. Runoff from the CBM (1,400,000 ML/y) includes greater areas of agricultural and grazing activity. Further event water quality data collected in each of these specific landuses and targeting event runoff will enhance the ability of building relationships to validate model predictions.

In relation to the overall load calibration results [\(Table 13\)](#page-42-0) the results suggest that TN and TP estimates generated better Moriasi statistics than TSS. This could be partly due to the fact that the Moriasi (2015) assessment criteria require TSS estimates to be closer to the observed load than nutrients when determining the performance rating.

7.2.2 CONTRIBUTION BY LAND USE AND SOURCE

The contribution of sediment and nutrients by land use are similar to the findings for previous models (Waters, 2006 and 2008). Similar results were found for the Fitzroy catchment (McCloskey, 2017) with comparative load contributions generally following landuse, grazing followed by cropping. Fitzroy is used for comparison to the QMDB due to its relatively similar range of land uses, rainfall and area.

Interestingly in the calibration of the loads along the Condamine catchment into the Balonne, the calibration for TP and TN align well in the upper part of the Condamine catchment at Warwick and Loudon Bridge gauging stations, but dramatically underestimate the load further downstream at Cotswold and Weribone calibration sites. Further investigation should be undertaken to identify why the model is underestimating in this area, or event samples to bolster the TN and TP vs discharge relationship used to create the observed flow. This may require further data collection for the lower end of the Balonne system, including gauging stations observing flow in this 300km stretch of river.

7.2.3 COMPARISON WITH GREAT BARRIER REEF PADDOCK TO REEF MODELLING

Comparing the GBR WQ models (McCloskey, 2017) to the QMDB modelled output allows some context be given to the model outputs. [Table 36](#page-97-0) shows the comparison of Constituent export loads to average annual flow and area of the catchments. For example when comparing the Fitzroy to Warrego, Paroo and Bulloo TSS export per unit area are similar whilst loads exported from the Balonne are approximately a quarter of the load per unit area of the Fitzroy basin. The other notable factor is that QMDB flows are less than a quarter of the annual flows of the GBR basins.

7.2.4 INDEPENDENT DATA VALIDATION OVER AN EVENT

The model predictions are encouraging when comparing sediment and nutrient concentrations from the model to an independent set of water quality samples collected at one gauging station in the Border Rivers at Mungindi [\(Figure 35](#page-51-0) and [Figure 36\)](#page-52-0). The modelled concentrations were within 100% of the measured concentrations which is extremely encouraging considering the hillslope/gully and streambank models were all originally developed as an average annual erosion model and were never intended to be used at a daily timestep. The TSS modelled concentrations aligned well with the grab samples for higher flows and underestimated the results at low flows. Given the majority of the sediment is transported in higher flows this allows for a good approximation by the model for TSS loads.

In addition, given TSS is aligning well at high flows and TN &TP correlate well to TSS [\(Table 10\)](#page-34-0), suggests that the modelling approach used to estimate TN and TP may require further investigation. The model results are very encouraging and further investigation is warranted into the input data behind the particulate nutrient generation models at high flows. It has been previously reported by McCloskey et al. (2017) that the subsurface nutrient data sets used as an input data set to the model may be too low.

8 CONCLUSIONS

The project was successful in collating the range of water quality data sets collected across the basins and deriving flow and concentration relationships. The data was then used to derive the most up to date estimates of catchment loads. A catchment water quality model was successfully built.

The model has been used to estimate constituent loads for the QMDB drawing on the latest modelling methods. This model provides improved estimates of sediment and nutrient loads compared to the previous model estimates in particular due to the advances made in hydrology calibration approach and the representation of the dominant erosion process namely hillslope, gully and streambank erosion. The method described in this report provides an alternative approach of deriving load estimates to calibrate a water quality model in areas where there is limited water quality data.

The modelling exercise has also provided an additional benefit to assist in the development of water quality guidelines by the Queensland Government in parts of the basin where data is not available.

This method provides a way forward to model sediment loads in data poor catchments. With further refinement and additional event data collection, the model can be used to inform where on ground works should be implemented and to priorities future research.

9 FUTURE RECOMMENDATIONS

Future effort to improve the model should focus on regionalisation of the constituent generation parameters. For example delivery ratios could be derived regionally based on local soil attributes. Similarly for gully and stream bank geometry. With limited desktop analysis regionalised parameters could be derived to improve spatial representation of sediment and nutrient generation processes.

The flow to TSS/TN/TP relationships can be significantly improved by assessing existing data collected at gauging stations and targeting specific flow ranges where data is currently limited at each of the calibration sites. This will do two things to improve the relationship. Firstly, allow for a better understanding of the loads generated in high flows versus low flows as well as improve the understanding of how concentrations vary on rising and falling limbs of the hydrograph at a site. Secondly, targeting landuse specific areas such as upland grazing and intensive cropping areas along the lower Condamine River will greatly enhance the spatial calibration of loads for specific industries.

10 REFERENCES

- Bartley, R, Speirs, WJ, Ellis, TW, Waters, DK 2012, A review of sediment and nutrient concentration data from Australia for use in catchment water quality models, *Marine Pollution Bulletin* 65 (4–9), 101–16
- Chiew, FHS, Siriwardena, L 2005, 'Estimation of SIMHYD parameter values for application in ungauged catchments' in *MODSIM 2005 International Congress on Modelling and Simulation: Advances and applications for management and decision making*, A Zerger & RM Argent (eds), Modelling and Simulation Society of Australia and New Zealand, Melbourne, December 2005, pp. 2883–89.
- (DNRM) Department of Natural Resources and Mines (2003), (includes Bulloo, Warrego, Paroo) River System Hydrology Report- Volume 1. Internal Report
- Doherty, J (2009), *PEST: Model-Independent Parameter Estimation, User Manual: 5th Edition*, Watermark Numerical Computing, Brisbane.
- DPI (1978). Western Arid Region Land Use Study Part IV. Queensland Department of Primary Industries. Technical Bulletin No. 23
- Ellis, R, Doherty, J & Searle, R (2009), "Applying parameter estimation and prediction uncertainty analysis to WaterCAST water quality models", *18th World IMACS Congress and MODSIM09 International Congress on Modelling and Simulation: Interfacing Modelling and Simulation with Mathematical and Computational Sciences*, eds. R.S. Anderssen, R.D. Braddock & L.T.H. Newham, Modelling and Simulation Society of Australia and New Zealand and International Association for Mathematics, Cairns, July.
- Ellis, R, Searle, R (2014), *Dynamic SedNet component model reference guide*, Queensland Department of Science, Information Technology, Innovation and the Arts, Bundaberg, Queensland.
- eWater Ltd 2012*, Source Catchments product information,* URL: www.eWater.com.au/products/ewatersource/for-catchments/, viewed May 2011.
- eWater Ltd 2013, *Source scientific reference guide (v3.5.0)*, eWater Ltd, Canberra.
- Li, Y., Migliacio, K., (2010) Water Quality Concepts, Sampling and Analyses. CRC Press (ISBN 9781420092677)
- Lyne, V. & Hollick, M. 1979, "Stochastic time variable rainfall-runoff modelling", Proceedings of the Hydrology and Water Resources Symposium, Perth, 10-12 September, Institution of Engineers National Conference Publication, No. 79/10, pp. 89-92
- McCloskey, G.L., Waters. D., Baheerathan, R., Darr, S., Dougall, C., Ellis, R., Fentie, B., Hateley, L. (2017). Modelling pollutant load changes due to improved management practices in the Great Barrier Reef catchments: updated methodology and results – Technical Report for Reef Report Cards 2015, Queensland Department of Natural Resources and Mines, Brisbane, Queensland.
- Moriasi, D. N., Arnold, J. G., Van Liew, M. W., Bingner, R. L., Harmel, R. D., Veith, T. L. (2007) Model Evaluation Guidelines for Systematic Quantification of Accuracy in Watershed Simulations. Transactions of the ASABE, Vol. 50(3) 885- 900.
- Moriasi, D. N., Gitau, M. W., Pai, N., Daggupati, P., (2015) Hydrologic and Water Quality Models: Performance Measures and Evaluation Criteria, *Natural Resources & Environmental Systems Community of ASABE*, Vol. 58(6): 1763-1785
- MDBA, (2017) What's in the Basin Plan? <https://www.mdba.gov.au/basin-plan/whats-basin-plan> accessed 31/10/2017

Qld Government (2017) *Soils data*,<https://www.qld.gov.au/environment/land/soil/soil-data> accessed 2/10/2017

Qld Department of Natural Resources and Mines (2016) *Queensland Government Water Monitoring Information Portal*, <https://water-monitoring.information.qld.gov.au/> accessed 2/1/2016

- QLUMP Queensland Department of Natural Resources and Mines website (accessed 22nd January, 2016) Queensland Land Use Mapping Program Background, Department of Natural Resources and Mine [https://www.qld.gov.au/environment/land/vegetation/mapping/qlump-background/#](https://www.qld.gov.au/environment/land/vegetation/mapping/qlump-background/)
- Renard, KG, Ferreira, VA 1993, RUSLE model description and database sensitivity, *Journal of Environmental Quality* 22 (3), 458–66.
- Renard, KG, Foster, GA, Weiss, DK, McCool, DK, Yoder, DC 1997, *Predicting soil erosion by water: a guide to conservation planning with the Revised Universal Soil Loss Equation*, United States Department of Agriculture, Agriculture handbook no 73, 404 pp.
- Rohde, K, Masters, B, Fries, N, Noble, R, Carroll, C 2008, *Fresh and marine water quality in the Mackay Whitsunday region 2004/05 to 2006/07*, Queensland Department of Natural Resources and Water for the Mackay Whitsunday Natural Resource Management Group, Australia.
- Silo Queensland Department of Natural Resources. (2004). The SILO Data Drill. Department of NRM. Available from http://www.nrm.qld.gov.au/silo/datadrill/datadrill_frameset.html
- Thornton, CM, Cowie, BA, Freebairn, DM, Playford, CL 2007, The Brigalow Catchment study. II. Clearing brigalow (Acacia harpophylla) for cropping or pasture increases runoff, *Australian Journal of Soil Research* 45 (7), 496–511.
- Turner, R, Huggins, R, Wallace, R, Smith, R, Vardy, S, Warne, MSJ 2012, *Sediment, nutrient and pesticide loads: Great Barrier Reef Catchment loads monitoring 2009–2010*, Department of Science, Information Technology, Innovation and the Arts, Brisbane.
- Waters, DK, Carroll, C, Ellis, R, Hateley, L, McCloskey, GL, Packett, R, Dougall, C, Fentie, (2014), *Modelling reductions of pollutant loads due to improved management practices in the Great Barrier Reef catchments – Whole of GBR, Technical Report, Volume 1*, Queensland Department of Natural Resources and Mines, Toowoomba, Queensland (ISBN: 978-1-7423-0999).
- Waters, D. (2006). Application of the EMSS water quality model for the Queensland Murray Darling catchment -Assessing the impacts of on-ground works. Technical Report of the Water Quality State-level Investment Project. Brisbane: Queensland Government - National Action Plan for Salinity and Water Quality.
- Waters, D. (2008) Water Quality monitoring and E2 modelling in the South West NRM region, Queensland Technical Report for South West NRM
- Waters, DK, Carroll, C, Ellis, R, Hateley, L, McCloskey, GL, Packett, R, Dougall, C, Fentie, (2014), *Modelling reductions of pollutant loads due to improved management practices in the Great Barrier Reef catchments - Whole of GBR, Technical Report, Volume 1*, Queensland Department of Natural Resources and Mines, Toowoomba, Queensland (ISBN: 978-1-7423-0999).
- Waters, D, Packett, R 2007, 'Sediment and nutrient generation rates for Queensland rural catchments—an event monitoring program to improve water quality modelling', in *Proceedings of the 5th Australian Stream Management Conference, Australian rivers: making a difference*, AL Wilson, RL Dehaan, RJ Watts, KJ Page, KH Bowmer & A Curtis (eds), Charles Sturt University, Thurgoona, NSW, 21–25 May 2007.
- Young, WJ, Prosser, IP & Hughes, AO, 2001, *Modelling nutrient loads in large-scale river networks for the National Land and Water Resources Audit,* CSIRO, Canberra.
- Yee Yet, J, Silburn, DM 2003, *Deep drainage estimates under a range of land uses in the QMDB using water balance modelling*, Queensland Department of Natural Resources and Mines, Toowoomba, Queensland QNRM03021.

APPENDICES

PARAMETERS ADOPTED IN SACRAMENTO MODEL

TABLE 16 - SACRAMENTO PARAMETERS FOR MOONIE AND BORDER RIVERS

TABLE 17 - SACRAMENTO PARAMETERS FOR SWNRM

TABLE 18 - SACRAMENTO PARAMETERS FOR CONDAMINE BALONNE MARANOA

OVERVIEW OF FUNCTIONAL UNIT CREATION

11.2.1 Condamine Balonne

Land use layer used: Qld Land use Mapping (SIR layer – Current) (Areas covered are based on 2012/2013 imagery - se[e Figure 37\)](#page-65-0)

Foliage Projected Cover - 2013

Rationale:

A diverse range of land use within the catchments allows for the use of the QLUM to group a range of Functional Units. The exception being grazing, which was split into two further groups based on the Foliage Projected Cover (FPC) – open grazing (less than 20% FPC) and forested grazing (>=20% FPC).

Categories:

As set out i[n Table 19.](#page-64-0)

TABLE 19 - FUNCTIONAL UNIT CATEGORIES FOR CONDAMINE BALONNE MODEL

FIGURE 37 - QLUMP CURRENCY BASED ON IMAGERY USED (DNRM QLUMP, 2016)

11.2.2 Moonie and Border Rivers

Land use layer used: 2006 Australian land use classification

Rationale:

As these two catchments both have sections of the area within NSW, which is not covered by Queensland Land use mapping, the Australian Land Use Mapping (ALUM) was used to create a complete coverage for this model. The Qld Land Use Mapping (QLUM) uses the standard set up by the ALUM, but extends it through mapping done by local operators at a finer scale.

Categories:

Used the same functional units as set for the Condamine Balonne Model applied to the ALUM as per [Table](#page-66-0) [20.](#page-66-0)

TABLE 20 - CRITERIA FOR FUNCTIONAL UNITS FOR MOONIE AND BORDER RIVER MODELS

11.2.3 SOUTH WEST NRM CATCHMENTS **Land use layer used:**

Qld Land use Mapping (SIR layer – Current) (Areas covered are based on 2006 imagery - see [Figure 37\)](#page-65-0)

Australian Land use mapping (small areas overlap with NSW)

Grazing Land Management Layer version 2 (SIR layer)

Rationale:

Grazing covers over 85% of the catchments of Bulloo, Paroo, Warrego and Nebine, using the Qld Land use Mapping. Land use was a surrogate of the landscape response to environmental factors grouping both natural and man-made processes. As within the 2008 E2 Model process (Waters, 2008) a broader definition of the Grazing areas within the South West Catchments was adopted. This was after consultation with SWNRM to understand how the landscape was broken up. A decision was made to adopt the broad Grazing Land Management Land Type Groups which were used in the State Rural Leasehold Land process in understanding the condition of land types. These categories were created to fill the grazing areas based on the largest types

For the non-grazing areas a similar grouping were made with QLUMP.

Categories:

Use the same categories as set out in **[Table 21](#page-68-0)**.

11.3 OVERVIEW OF SUBCATCHMENT CREATION

The sub catchment generation was undertaken for the project in three phases.

Phase 1: Discussion with end users

Meeting were arranged with the three NRM bodies to ascertain the likely uses of the model and also the required planning level to assist the planning work within the region. This discussion helped identify the appropriate size for the sub catchment size.

Phase 2: Source Catchment DEM catchments generation

Using the AUS STRM 1 SEC DEM –Hydrological enforced version 1 the area surrounding catchments were extracted along with a 25km buffer. This was used for the basis of the creation of sub catchments within the program Source Catchments. Sub catchment layers were generated using individual clipped DEM for the individual catchment due to the size of the data that needed to be processed (>1Gb) using the Geographic Scenario wizard process. This was an iterative process on a variety of sub catchment sizes to ensure all of the Gauging Stations and Storages were identified as nodes. Draft sub catchments were exported at a variety of sub catchment sizes.

Phase 3: Final Sub catchment generation

Utilizing the understanding gained through a variety of sub catchment sizes, the final set of sub catchment were generated through combining draft sub catchment of different sizes to ensure all required nodes are correctly represent in the sub catchment model. The table below outlines some of these modifications

The final associated sub catchment nodes and links models were modified to allow for correct flow direction on reloading back into Source.

CALIBRATION RESULTS

TABLE 22 - FULL CALIBRATION RESULTS FOR ALL GAUGES USED IN THE HYDROLOGICAL CALIBRATION PROCESS

EMC/DWC VALUES USED

TABLE 23 - MEAN EVENT CONCENTRATION USED FOR THE MODEL BY FUNCTIONAL UNIT

SWNRM

* EMC (Mean Event Concentration) represents the hillslope generation. The delivery ratio and gully model is applied separately.

ANNUAL LOADS AND FLOW

TABLE 24 – SOUTHWEST NRM MODELLED ANNUAL LOADS AND FLOW

Table 25 – BORDER RIVERS MODELLED ANNUAL LOADS AND FLOW

 $\mathcal{L}^{\mathcal{L}}$

Table 26 – CONDAMINE BALONNE AND MARANOA ANNUAL LOADS AND FLOW

Balonne Region (includes cond. + Maranoa)

LOAD VALIDATION RESULTS

TABLE 27 – PERCENTAGE OF SITES THAT MEET THE GIVEN MORIASI STATISITICS

TABLE 28 – YEARLY TSS CALIBRATION RESULTS

TABLE 29 - MONTHLY TSS CALIBRATION RESULTS

TABLE 30 - YEARLY TN CALIBRATION RESULTS

TABLE 31 - MONTHLY TN CALIBRATION RESULTS

TABLE 32 - YEARLY TP CALIBRATION RESULTS

TABLE 33 - MONTHLY TP CALIBRATION RESULTS

NITROGEN AND PHOSPHORUS RELATIONSHIP FIGURES WITH FLOW

FIGURE 38 - SWNRM LOG DISCHARGE VS LOG TN/TP RELATIONSHIPS

FIGURE 39 - SWNRM LOG TSS VS TN/TP RELATIONSHIPS

FIGURE 40 - CB LOG DISCHARGE VS LOG TN/TP RELATIONSHIPS

FIGURE 41 - CB LOG TSS VS TN/TP RELATIONSHIPS

FIGURE 42 - BDMN LOG DISCHARGE VS LOG TN/TP RELATIONSHIPS

FIGURE 43 - BDMN LOG TSS VS TN/TP RELATIONSHIPS

COMMUNICATION WITH QMDB STAKEHOLDERS

TABLE 34 - COMMUNICATION WITH MAJOR STAKEHOLDERS IN THE QMDB

TABLE 35 – LANDUSE CONTRIBUTION TO EXPORT PER UNIT AREA EXPRESSED AS A PERCENTAGE OF TOTAL

TABLE 36 - COMPARISON OF MDB LOADS TO GBR LOADS BY BASIN

